

1
2
3
4 **Sperm exposure to carbon-based nanomaterials causes abnormalities in early**
5
6
7 **development of purple sea urchin (*Paracentrotus lividus*)**
8
9

10
11
12
13
14 **Tina Mesarič^a, Kristina Sepčič^a, Damjana Drobne^{a,b}, Darko Makovec^{b,c}, Marco Faimali^d,**
15
16 **Silvia Morgana^d, Carla Falugi^e, Chiara Gambardella^{d*}**
17
18
19
20
21
22

23 ^a Department of Biology, Biotechnical Faculty, University of Ljubljana, Slovenia
24

25 ^b Centre of Excellence in Nanoscience and Nanotechnology, Ljubljana, Slovenia
26

27 ^c Institute Jožef Stefan, Jamova 39, Ljubljana, Slovenia
28

29 ^d Institute of Marine Sciences, National Research Council, Genova, Italy
30
31

32 ^e Department of Environmental and Life Sciences, Università Politecnica delle Marche, Ancona,
33
34 Italy
35
36
37
38
39
40
41

42 ***Corresponding author: Chiara Gambardella**
43

44 ISMAR-CNR
45

46 Via de Marini 6
47
48

49 16149 Genova, Italy
50

51
52 Tel: +39-010-6475429
53

54 Fax: +39-010-6475400
55

56
57 E-mail: chiagamba@gmail.com, chiara.gambardella@ge.ismar.cnr.it
58
59
60
61
62
63
64
65

1
2
3
4 **Abstract**
5
6
7
8

9 We examined egg fertilisation in purple sea urchin (*Paracentrotus lividus*) after sperm exposure
10 to carbon-based nanomaterials, carbon black (CB) and graphene oxide (GO), from 0.0001 mg/L
11 to 1.0 mg/L. Gastrula stage embryos were investigated for acetylcholinesterase and
12 propionylcholinesterase activities, and their morphological characteristics. Plutei were analysed
13 for morphological abnormalities, with emphasis on skeletal rod formation. Egg fertilisation was
14 significantly affected by CB, at all concentrations tested. Loss of cell adhesion at the gastrula
15 surface was observed in eggs fertilised with sperm treated with CB. However, concentration-
16 dependant morphological anomalies were observed in the gastrulae and plutei formed after sperm
17 exposure to either CB or GO. The activities of both cholinesterases decreased in the gastrulae,
18 although not in a concentration-dependent manner. These effects appear to arise from physical
19 interactions between these carbon-based nanomaterials and the sperm, whereby nanomaterials
20 attached to the sperm surface interfere with fertilisation, which leads to disturbances in the
21 signalling pathways of early embryonic development. Reduced cholinesterase activity in
22 gastrulae from eggs fertilised with nanomaterial-treated sperm confirms involvement of the
23 cholinergic system in early sea urchin development, including skeletogenesis.
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44

45
46
47 **Keywords:** acetylcholinesterase, carbon-based nanomaterials, development,
48 propionylcholinesterase, sea urchin, sperm.
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

1. Introduction

Naturally occurring and industrially derived carbon-based nanomaterials are the most abundant nanomaterials in the environment (Hussain et al., 2009). Carbon black (CB) is a form of amorphous carbon that has extensive use in industrial applications (e.g., rubber products, paints, plastic, inks). CB is manufactured by controlled vapour-phase pyrolysis of hydrocarbons (CAS RN, 2013; Sorahan and Harrington, 2007), and it should not be confused with ‘black carbon’, which is a product of incomplete combustion of fossil fuels and biomass (Goldberg, 1985; Masiello and Druffel, 1998; Middleburg et al., 1999). Graphene oxide (GO) is formed of monomolecular sheets in which the carbon atoms are arranged in a graphene-like honeycomb lattice that is interrupted and terminated with numerous oxygen-containing functional groups (e.g., epoxy, carboxyl, hydroxyl). GO is an advanced and highly investigated nanomaterial that has physical and chemical characteristics that make it attractive for a variety of commercial and industrial purposes (Rao et al., 2009; Wang et al., 2011).

Due to the hydrophobic nature of carbon-based nanomaterials and their high surface-adsorption potential (Xia et al., 2011; Mesarič et al., 2013a), they can attach to different surfaces, including organisms, which might present a serious risk for biota (Klaine et al., 2008; Mesarič et al., 2013b). Although in recent years several ecotoxicological studies, on the impact of nanomaterials on marine organisms, have been performed (Canesi et al., 2008; Nielsen et al., 2008; Canesi et al., 2010a, 2010b; Rajasree et al., 2010; Miglietta et al., 2011; Carata et al., 2012; Falugi et al., 2012; Matranga et al., 2012; Miller et al., 2012; Ates et al., 2013; Mesarič et al., 2013b; Gambardella et al., 2014), there remains a lack of knowledge in this field.

The Mediterranean purple sea urchin (*Paracentrotus lividus*) is frequently used as a model in marine toxicology studies. **This species is widely diffused in benthic littoral zone of**

1
2
3
4 **the Mediterranean and Atlantic (Boudouresque and Verlaque, 2001), an area that is**
5 **strongly affected by human activities and pollution (ASTM, 2004). In these areas, *P. lividus***
6 **represents a key species, able to remodel hard bottoms and to cause barren grounds in**
7 ***Posidonia* prairies by its grazing activity. Thus, this species is relevant for the control and**
8 **organization of benthonic communities. While adult specimens are relatively resistant to**
9 **pollution, the naked sperm and larval stages are directly affected by the presence of**
10 **contaminants.** Recently, it was reported that *P. lividus* planktonic larvae and benthonic adults
11 can be affected by nanomaterials. Metal oxide nanomaterials (e.g, SnO₂, CeO₂, Fe₃O₄) have been
12 found inside *P. lividus* immune cells after forced ingestion (Falugi et al., 2012). Furthermore,
13 metal and metal oxide nanomaterials were reported to cause significant developmental anomalies
14 in *P. lividus* embryos and larvae (Gambardella et al., 2013; Manzo et al., 2013). Similar data
15 were reported by Manno et al. (2013) for carbon-based nanomaterials, which were shown to
16 accelerate *P. lividus* embryonic development and induce embryonic malformations. In another
17 study by the same group, carbon-based nanomaterials were reported to alter the expression of
18 genes involved in biomineralisation (Manno et al., 2012). Carbon-based nanomaterials have also
19 been shown to affect gene expression and development of sea urchin larvae, where they caused
20 asymmetrical cell division and altered gene expression, again, with acceleration of development
21 (Matranga et al., 2012). Gambardella et al. (2013) also provided evidence of altered embryonic
22 development of sea urchins, along with effects of cholinergic signalling system in embryos, after
23 fertilisation of eggs with sperm previously exposed to metal and metal oxide nanomaterials.
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51

52
53 Neurotransmitters such as acetylcholine (ACh), biogenic amines, and γ -aminobutyric
54 acid, are functionally active throughout ontogenesis (Buznikov et al., 1996). In particular, the
55 cholinergic system is involved in both neuronal and non-neuronal signal transduction (Buznikov
56
57
58
59
60
61
62
63
64
65

1
2
3
4 et al., 1996), and its non-neuronal signalling function is best studied in processes of fertilisation
5
6 and development in different animal species. Falugi et al. (1993) demonstrated the dependence of
7
8 sperm motility on ACh receptors, and Baccetti et al. (1995) and Angelini et al. (2003) reported
9
10 that sperm–egg fusion is based on signalling molecules of the cholinergic system. ACh
11
12 availability and the correct functioning of both nicotinic and muscarinic ACh receptors, are
13
14 responsible for ionic and cytoskeletal dynamics that are crucial for correct embryonic
15
16 development of *P. lividus* (Aluigi et al., 2010; Falugi et al., 2008a; Falugi and Aluigi, 2012;
17
18 Ravera et al., 2006). As ion channels gated by ACh, nicotinic receptors allow the inward passage
19
20 of Na⁺, while upon activation by ACh, muscarinic receptors trigger an intracellular transduction
21
22 cascade that results in the release of Ca²⁺ from intracellular stores. It is expected that ACh
23
24 receptor agonists and antagonists can either have direct effects on the ACh receptors, or have
25
26 indirect effects, such as effects on the enzyme acetylcholinesterase (AChE), which degrades
27
28 ACh. Thus, both ACh agonists and antagonists can affect the level of receptor activation, which
29
30 can lead to alterations in signal transduction and cell-to-cell communication.
31
32
33
34
35
36
37

38 In the present study, we investigated fertilisation and early development of *P. lividus*
39
40 when the eggs were fertilised with sperm that had been incubated in CB or GO suspensions. We
41
42 assessed the following biomarkers for the effects of exposure of sperm to these carbon-based
43
44 nanomaterials: (i) egg fertilisation; (ii) morphological features of embryos and larvae (i.e.,
45
46 gastrula and pluteus stages); and (iii) activities of two embryo (i.e., gastrula) cholinesterases:
47
48 AChE (E.C. 3.1.1.7) and propionylcholinesterase (PChE; E.C. 3.1.1.8).
49
50
51
52

53 We expected here that these carbon-based nanomaterials will affect egg fertilisation by
54
55 the previously exposed sperm, due to the high adsorption potential of these nanomaterials.
56
57 Furthermore, these exposed sperm might transport the carbon-based nanomaterials to the eggs,
58
59 where they can then interfere with fertilisation events. As the cholinergic system is involved in
60
61
62
63
64
65

1
2
3
4 regulation of the early embryonic developmental events in *P. lividus* (Angelini et al., 2003;
5
6 Falugi et al., 2008a; Falugi and Aluigi, 2012), we hypothesised that any developmental anomalies
7
8 would also be accompanied by altered cholinesterase activities in the developing embryos.
9

10 11 12 13 14 **2. Materials and methods**

15 16 17 18 19 **2.1. Nanomaterials**

20
21 The CB powder was from PlasmaChem GmbH (Germany). The mean size of the primary
22
23 particles was ~20 nm (see Supplementary Information, **Fig. S1**). The GO was from Graphene
24
25 Supermarket, as dry flakes with a mean size of 0.5 μm to 5.0 μm . At least 80% of the GO was of
26
27 single-layer thickness, and according to the producer specification, the GO contained 20% by
28
29 weight of oxygen. The CB and GO suspensions (final stock concentration, 1 mg/mL) were
30
31 freshly prepared in 0.22- μm -filtered natural sea water (FNSW). The CB and GO stock
32
33 suspensions were sonicated for 30 min in an ultrasonic bath (Falc sonicator, model LBS1; Italy),
34
35 as described in Canesi et al. (2010a). **The assessment of the nanomaterial size of the CB and**
36
37 **GO suspensions at a final concentration of 1 mg/mL by using dynamic light scattering was**
38
39 **not possible as these nanomaterials strongly flocculate in FNSW (Mesarič et al., 2013b;**
40
41 **Mesarič et al., 2015).** However, their characterisation in distilled water (Mesarič et al., 2013a)
42
43 showed relatively high negative ζ -potentials (from -32 to -36 mV) for both of these
44
45 nanomaterials. Furthermore, dynamic light scattering measurements of the CB suspension in
46
47 distilled water showed substantial agglomeration of the particles, with sizes ranging from 60 nm
48
49 to ~300 nm (Mesarič et al., 2013b). For the GO suspension in distilled water, analysis using
50
51 dynamic light scattering was not possible because of the extremely anisotropic shapes of the
52
53 primary particles (see Supplementary Information, **Fig. S2**). **The ξ -potential of nanomaterials**
54
55
56
57
58
59
60
61
62
63
64
65

1
2
3
4 suspended in the FNSW was determined to be much lower (around -14 mV) to that in the
5
6 distilled water because of electrostatic screening effects related to the very high ionic
7
8 strength (Mesarič et al., 2015).
9

10
11 Metals are used during nanomaterial synthesis, and the presence of metal ions (mainly Cu,
12
13 Gd, Zn, Al and Fe) in nanomaterial leachates can induce significant toxicity in aquatic
14
15 invertebrates and vertebrates. Thus, the presence of these metal ions should be considered
16
17 when evaluating the toxicity of carbon-based nanomaterials (Hull et al., 2009). In line with
18
19 this, FNSW suspensions of CB or GO used in tests described in this work, and also in
20
21 experiments described in Mesarič et al. (2015), were analysed using energy-dispersive X-ray
22
23 spectroscopy (EDS). These analyses revealed the absence of impurities in the form of metal
24
25 ions. In both carbon-based nanomaterial suspensions in FNSW, also some evident cubic
26
27 crystals were identified. Using EDS, they were confirmed to be composed of seawater salts
28
29 (Fig. S3).
30
31
32
33
34
35
36
37

38 2.2. Egg fertilisation

39
40 Adult samples of *P. lividus* were collected along the coast of Genoa (Italy). They were
41
42 transported to the laboratory in a refrigerated bag that was wrapped in towels that had been
43
44 previously wet with sea water (Amemiya, 1996).
45
46
47

48 The spawning of the gametes was triggered by oral administration of 1‰ ACh in FNSW,
49
50 with the eggs collected in standard FNSW. The sperm were collected directly from the genital
51
52 pores and maintained in aliquots of 200 µL at 4 °C. The sperm from three different specimens
53
54 were mixed. The experiments were repeated three times during the breeding season, with each
55
56 carried out in triplicate. **Ten microliter aliquots of sperm suspensions were exposed to 1 mL**
57
58 **CB or GO nanomaterial suspensions. Such a small volume of sperm was used since it has**
59
60
61
62
63
64
65

1
2
3
4 **been demonstrated by contact with seawater within a brief period that the motility is**
5
6 **initiated for almost 100% of sea urchin spermatozoa (Gibbons, 1981). CB and GO**
7
8 **nanomaterial suspensions were prepared at the serial concentrations of: 0.0001, 0.001, 0.01,**
9
10 **0.1 and 1.0 mg/L, according to Gambardella et al. (2013, 2015). Since to date no regulatory**
11
12 **guidelines are available for nanomaterials, the selected concentrations were chosen on the**
13
14 **basis of those used for the reference toxicant (copper nitrate) for *P. lividus* spermotoxicity**
15
16 **test reported by the Italian guidelines (Arizzi Novelli et al., 2001, 2007a, b). No**
17
18 **agglomerates of nanomaterials were observed under microscope in the FNSW, which is**
19
20 **probably due to the short-term exposure period (1 h), as well as to low concentration of**
21
22 **nanomaterials tested.** Controls were performed by adding 10 μ L standard FNSW instead of CB
23
24 or GO nanomaterial suspensions. After 1 h of incubation of the sperm with the CB or GO
25
26 nanomaterials, FNSW was added to each incubation mixture to a 1-mL final volume, and the
27
28 suspensions **were centrifuged at 735 x g for 3 min at 4 °C (Mod. 5415 D, Eppendorf). This**
29
30 **experimental basis was used since at this speed sperm cells survive and are able to fertilize**
31
32 **the eggs, whereas their motility could be affected at higher speed centrifugation forces**
33
34 **(Tash et al., 2001).** The supernatants were discarded to remove the free nanomaterials, and the
35
36 sediments (sperm with any bound nanomaterials) were resuspended in 1 mL FNSW. Ten
37
38 microliters of these supernatants were then added to multiwell capsules containing 10 mL FNSW
39
40 with approximately 300 eggs/mL. Under normal conditions, the egg activation occurred within
41
42 60 s of the sperm addition **(Alberts et al., 1989).** To minimise the possibility of direct effects of
43
44 the nanomaterials on the eggs and zygotes that might occur when the original protocol was used
45
46 (Ghirardini et al., 2005), rinsing of the sperm was introduced before the fertilisation, as described
47
48 above.
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

1
2
3
4 To determine the egg fertilisation in these incubations, 2 mL FNSW (containing
5
6 approximately 600 fertilised and unfertilised eggs) was collected from each well 20 min after
7
8 sperm addition, and fixed with 4% paraformaldehyde in FNSW for several hours. The eggs were
9
10 then rinsed in physiological solution (Tyrode, 1910). The samples were observed under light
11
12 microscopy (DM3000B, Leica, Germany) at 40× magnification, and the percentages of eggs
13
14 showing a swollen perivitelline space, as the indication of fertilisation, were determined by
15
16 counting under the microscope, from random fields with up to 600 eggs counted per treatment.
17
18
19
20

21 **The reliability of the test was verified using the reference toxicant copper nitrate,**
22
23 **according to Arizzi Novelli et al. (2002). The acceptability of the results was fixed at a**
24
25 **fertilization rate > 70% in control tests (Arizzi Novelli et al., 2001; Volpi Ghirardini and**
26
27 **Arizzi Novelli, 2001).**
28
29
30
31
32

33 **2.3. Analyses of embryotoxicity and skeletogenesis**

34
35 The remaining 8 mL FNSW that contained ~2400 fertilised and unfertilised eggs (see section 2.2)
36
37 were left to develop at 18 °C in a thermostatic room, and were sampled at 24, 48 and 72 h, which
38
39 corresponded to the gastrula, early pluteus and pluteus stages, respectively. Unfertilised eggs and
40
41 dead embryos were discarded 15 h after fertilisation.
42
43
44

45 For each sample, after 24, 48 and 72 h, approximately 100 larvae were fixed in cold
46
47 methanol and 20% polyethylene glycol (Sigma, Italy). The developmental stages were monitored
48
49 under light microscopy (DM3000B, Leica, Germany), including the formation of skeletal rods,
50
51 and classified on the basis of morphology and synchronicity of development, as compared to the
52
53 controls. This classification was according to the specific anomalies identified and recorded by
54
55 Carballeira et al. (2012) and Gambardella et al. (2013). **The acceptability of the results was**
56
57
58
59
60
61
62
63
64
65

1
2
3
4 **fixed at a percentage of normal development >70% in control tests (Arizzi Novelli et al.,**
5
6 **2002).**
7
8
9

10 11 **2.4. Cholinesterase activity**

12
13
14 The AChE and PChE activities were measured by the Ellman method (Ellman et al., 1961),
15
16 modified for spectrophotometer use (6405 UV/VIS; Jenway). The 24 h old embryos at the
17
18 gastrula stage were collected and maintained for 2 weeks at -20 °C. Prior to the AChE and PChE
19
20 activity assays, these samples were homogenised, sonicated for 25 min in a bath sonicator (model
21
22 LBS1; FALC, Italy), passed through a fine syringe needle (Ultrafin 29G, 12.7 mm length) in the
23
24 presence of 1% Triton X100, **and centrifuged at 18,363 x g** for 30 s at 4 °C **(Mod. 5415 D,**
25
26 **Eppendorf)**. The kinetics of the AChE and PChE activities were measured by following
27
28 absorbance at 412 nm. The velocity of substrate cleavage was measured for 3 min, and compared
29
30 with the linear equation of a standard curve that had been previously obtained by supplying
31
32 known amounts of the cholinesterases. The protein content in the supernatants of the control and
33
34 exposed gastrulae was measured by the Bradford method (Biorad, USA), and referred to a
35
36 standard curve obtained with bovine serum albumin (Sigma, USA). The cholinesterase units were
37
38 calculated as enzyme activity/min/mg protein. The measurements were performed in triplicate,
39
40 using gastrulae from different parents, and were repeated three times during the reproductive
41
42 season of *P. lividus*.
43
44
45
46
47
48
49
50
51
52

53 **2.5. Statistical analysis**

54
55 Statistical analysis of the cholinesterase activities for determination of significant differences
56
57 between control and treated samples was performed using ANOVA (Bonferroni non-parametric
58
59 *post-hoc* tests). A *p* value lower than 0.05 (*p* <0.05) was considered significant. The data are
60
61
62
63
64
65

1
2
3
4 presented as means \pm standard error of three experiments, each of which was performed in
5
6 triplicate. The significant differences among the treatments and concentrations are reported in the
7
8 Figure legends.
9

10 11 12 13 14 **3. Results**

15 16 17 18 **3.1. Egg fertilisation by sperm exposed to the carbon-based nanomaterials**

19 The egg fertilisation by the *P. lividus* sperm was demonstrated by the cortical reaction of the egg
20
21 that occurs in response to the first sperm–egg contact. This can be visualised by the elevation of
22
23 the fertilisation layer and the appearance of the perivitelline space. The eggs showing this feature
24
25 were considered as fertilised, and were counted. The counts under the different treatments were
26
27 compared to the fertilisation of the control eggs, expressed as percentages. **In all experiments,**
28
29 **controls showed 96 \pm 3 % of fertilised eggs.** After the sperm were exposed to CB, the egg
30
31 fertilisation was decreased by about 50% in comparison with the control, at all of the
32
33 concentrations tested (0.0001-1.0 mg/L; **Fig. 1A**). In the case of the sperm exposed to GO, the
34
35 egg fertilisation was not affected (**Fig. 1B**). **The repeatability of the experiments was tested by**
36
37 **using copper nitrate as a reference toxicant. The test had a good repeatability, with an egg**
38
39 **fertilisation decrease of about 50% at 0.01 mg/L (not shown), similar to previous results**
40
41 **reported by Manzo et al. (2008).**
42
43
44
45
46
47
48
49
50
51
52

53 **3.2. Morphological characteristics of gastrulae and plutei**

54 The embryos sampled at the gastrula stage (24 h from fertilisation) that were obtained from eggs
55
56 fertilised with *P. lividus* sperm that had been exposed to these carbon-based nanomaterials
57
58
59
60
61
62
63
64
65

1
2
3
4 showed anomalous and arrested development (**Fig. 2A-D**). The lowest concentration of CB
5
6 (0.0001 mg/L) induced anomalous and arrested development in 3% of these embryos (**Fig. 2A**).
7
8 The proportion of anomalous embryos greatly increased with exposure to the increasing
9
10 concentrations of CB, and reached >80% at the highest concentration tested (1.0 mg/L; **Fig. 2A**).
11
12 Following this CB exposure of the sperm, in the 72-h-old larvae at the pluteus stage, the
13
14 anomalous plutei included developmentally delayed forms and those with skeletal damage (e.g.,
15
16 incomplete skeletal rods, crossed skeletal tips). These were seen in greater proportions than in the
17
18 controls (**Fig. 2B**). Here, the highest proportions of anomalous larvae were seen at the lowest CB
19
20 concentrations tested (87% and 80% at 0.0001 and 0.001 mg/L, respectively). At the highest CB
21
22 concentration tested (1 mg/L), there was 54% anomalous larvae. GO induced a concentration-
23
24 dependent increase in anomalies in the 24-h-old embryos (**Fig. 2C**). As for CB, non-developed
25
26 embryos did not represent a significant percentage at any of the concentrations tested (1%-2%).
27
28 The effects of GO on the development of the 72-h-old larvae was similar at all concentrations
29
30 tested, with an overall mean of 70% \pm 10% anomalies, as compared to the control (**Fig. 2D**).
31
32
33
34
35
36
37

38 **The morphology of the 24-h-old embryos at the gastrula stage obtained from the**
39 **eggs fertilised with *P. lividus* sperm that had been exposed to these carbon-based**
40 **nanomaterials is illustrated in Fig. 3 (upper panel). The embryos that developed from eggs**
41 **fertilised with sperm exposed to the increasing concentrations of CB showed deterioration**
42 **in their development. Anomalies in these embryos were observed at all of the CB**
43 **concentrations tested, and were represented by irregular embryo shapes due to an**
44 **anomalous migration of primary mesenchymal cells (Fig. 3a-f), by loss of cell-cell adhesion**
45 **at the gastrula surface (Fig. 3c-f), as shown by the break-down of the epithelium in single**
46 **separate cells (Fig. 3a-f), and by the lack of skeleton (Fig. 3 g). For the sperm exposure to**
47 **GO (in the concentration range from 0.001 to 1 mg/L), the 24-h-old embryos also showed**
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

1
2
3
4 **anomalies (Fig. 3, lower panels). The main anomalies were represented by the break-down**
5 **of the epithelium in single separate cells (Fig. 3b), by irregular embryo shapes (Fig. 3c-f),**
6 **and by the incomplete skeletogenesis process (Fig. 3g).**
7
8
9

10
11 Early pluteus larvae observed 48 h after fertilisation with *P. lividus* sperm that had been
12 exposed to these carbon-based nanomaterials showed abnormalities in the skeletal architecture,
13 and the larval shape and size, for both CB and GO exposure **at all tested concentrations (Fig.**
14 **4A)**. This trend was also followed in the 72-h-old larvae at the pluteus stage, which showed time-
15 dependent increases in the anomalies after egg fertilisation with the CB-exposed sperm, up to the
16 point where their development was completely arrested. At 72 h, these larvae had the shape of the
17 anomalous gastrula stage, they were smaller in size than the controls, and they lay at the bottom
18 of the wells, without swimming (**Fig. 4B, f**). The 72-h-old larvae obtained from eggs fertilised
19 with the GO-exposed sperm mainly showed anomalies in the skeletal architecture (e.g., lack of
20 arms, crossed apex tips, arms not oriented to the mouth).
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37

38 **3.3. Cholinesterase activities**

39
40 Significant differences were observed in the cholinesterase activities in the embryos after
41 fertilisation with the *P. lividus* sperm that had been exposed to these carbon-based nanomaterials.
42 The AChE activities were significantly lowered in 24-h-old embryos obtained after the sperm had
43 been exposed to either CB or GO, at all of the concentrations tested (**Fig. 5A, B**). The sperm
44 exposure to CB induced an overall mean of 35% inhibition of the AChE activity (**Fig. 5A**), and a
45 similar overall mean inhibition of 40% after the sperm exposure to GO (**Fig. 5B**). This inhibition
46 of the AChE activity was not dose-dependent. A similar trend was observed for the PChE activity
47 (**Fig. 5C, D**). In the embryos obtained from the sperm exposed to CB, the inhibition of the PChE
48 activity was also not dose-dependent, and showed an overall mean inhibition of 75% compared to
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

1
2
3
4 the control, across all of the concentrations tested (**Fig. 5C**). PChE activity was also inhibited in
5
6 the embryos that developed from the eggs fertilised with the sperm exposed to GO; however, at
7
8 the GO concentration of 0.001 mg/L, there was an increase in the enzyme activity (**Fig. 5D**).
9
10

11 12 13 14 **4. Discussion**

15 16 17 18 19 **4.1. Egg fertilisation by sperm exposed to CB and GO**

20
21 The data from the present study indicate that the *P. lividus* sperm exposure test appears to be very
22
23 sensitive to **carbon-based nanomaterials, such as CB and GO**. Here, exposure of the sperm to
24
25 even the lowest concentration of CB tested (0.0001 mg/L) induced adverse effects, while
26
27 exposure of the *P. lividus* embryos themselves to the same type of nanomaterial at 66 mg/L was
28
29 previously reported to be non-embryotoxic (Miglietta et al., 2011). **It was shown previously**
30
31 **that sea urchin sperm cells are more sensitive to toxic compounds than the**
32
33 **embryos (Manzo et al., 2008). Male gametes are released naked into the**
34
35 **seawater, exposing their membrane receptors directly to the environment and**
36
37 **therefore also to the contaminants (Falugi and Prestipino, 1987), including**
38
39 **nanomaterials. Conversely, embryos are surrounded by a fertilisation envelope,**
40
41 **that makes them more resistant to different environmental factors (Giudice,**
42
43 **1973).**
44
45
46
47
48
49

50
51 Sperm cells can be affected by different substances in a direct manner, with reduction or
52
53 suppression of egg fertilisation, and/or indirectly, where transmission of damage to the offspring
54
55 occurs. In the present study, we observed both types of effects in the case of the exposure of the
56
57 *P. lividus* sperm to CB, but only indirect effects when the sperm were exposed to GO. **We**
58
59 **believe that physical interactions, such as attachment of the carbon-based nanomaterials to**
60
61
62
63
64
65

1
2
3
4 **the sperm surface, will be one of the main reasons for the observed decrease in egg**
5 **fertilisation by these sperm that is caused by CB. It should be considered that the process of**
6 **low-speed and short-term centrifugation, that was used here to separate the free**
7 **nanomaterials from the sperm-bound nanomaterials (Chapter 2.2), might have also**
8 **enhanced the attachment of nanomaterials to the sperm surface. However, in such a case**
9 **one would expect that the fertilization ability of sperm exposed to GO, which forms much**
10 **larger particles than CB, would be also considerably affected.**
11
12
13
14
15
16
17
18
19
20

21 The adsorption potential of nanomaterials has already been suggested as being
22 responsible for their biological effects. For example, ZnO-induced sperm toxicity and
23 embryotoxicity in *P. lividus* was reported to be related not only to the zinc ions, but also to
24 surface interactions of particle/ aggregates with the target cells/ organisms (Manzo et al., 2013).
25
26 Similarly, CB has been reported to attach to sperm of the alga *Fucus serratus*, thus removing the
27 sperm from the suspension, which significantly affected the egg fertilisation by these sperm
28 (Nielsen et al., 2008).
29
30
31
32
33
34
35
36
37

38 These reports and the data from the present study are in line with the study of Xia et al.
39 (2011), who calculated the surface adsorption indices for different groups of nanomaterials, and
40 reported that carbon-based nanomaterials have considerably higher adsorption potential
41 compared to metal-based nanomaterials. These differences in adsorption potential appear to be
42 responsible for the observed lower adverse effects of metal and metal-based nanomaterials
43 (Gambardella et al., 2013), in comparison to the carbon-based nanomaterials investigated in the
44 present study under the same experimental set-up. Also, these data show different behaviours of
45 the two carbon-based nanomaterials tested, in terms of egg fertilisation by the exposed sea urchin
46 sperm: the exposure to GO did not affect the egg fertilisation, while the exposure to CB
47 significantly decreased it.
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

1
2
3
4 These different behaviours can be explained by the different characteristics of these two
5
6 carbon-based nanomaterials, such as their sizes, shapes, and surface adsorption potentials.
7
8 **Although similar ζ -potentials of the GO and CB suspensions were measured in FNSW**
9 **(Mesarič et al., 2015),** these two nanomaterials show different inherent hydrophilicities. CB is
10
11 inherently relatively hydrophobic; however, it adsorbs charged species (ions and charged
12
13 molecules) onto its surface from the liquid medium. On the contrary, GO has a hydrophilic
14
15 character due to its oxygen-containing functional groups (e.g., epoxy, carboxyl, hydroxyl). Xia et
16
17 al. (2011) also showed that pure carbon-based nanomaterials like CB show much greater surface
18
19 adsorption potential than polar carbon nanomaterials like GO. We believe that these properties of
20
21 CB allow stronger adsorption of this nanomaterial to the sperm cells, resulting in higher level of
22
23 disturbances during the egg fertilization process.
24
25
26
27
28
29
30
31
32

33 **4.2. Early developmental alterations**

34 35 36 37 38 **4.2.1. Inhibition of cholinesterase activity in embryos**

39
40 We found significant, but not dose-dependent, inhibition of the activities of both of the enzymes
41
42 investigated, as AChE and PChE, in gastrulae obtained after egg fertilisation with sperm exposed
43
44 to both CB and GO. **Considering that ACh is absent in sea urchin sperm up to the**
45
46 **fertilization process, this molecule, as well as its related enzymes (Angelini et al., 2004;**
47
48 **Falugi et al., 2008b), were not measured in the sperm in the present study. We measured**
49
50 **the AChE activity only in the embryos, since this enzyme plays multiple roles during**
51
52 **embryonic development (e.g. it is involved in cell-to-cell communication (Falugi and Aluigi,**
53
54 **2012), modulation of apoptosis (Zhang et al., 2002) and driving morphogenetic movements**
55
56 **(Drews, 1975) independently from the cholinergic system.**
57
58
59
60
61
62
63
64
65

1
2
3
4 As also reported previously (Falugi and Aluigi, 2012; Gambardella et al., 2013), greater effects
5
6 were seen here on the PChE activity. However, when sperm were exposed to metal (Ag, Co;
7
8 (Gambardella et al., 2013)) or metal oxide (TiO₂; Gambardella, personal communication)
9
10 nanomaterials, these effects were not as large as in the present study with the same experimental
11
12 set-up. Gambardella et al. (2013) reported both increased and decreased AChE and PChE
13
14 activities in *P. lividus* embryos.
15
16
17

18
19 The cholinergic system is active in the acrosome, neck region, and flagellum of the sea
20
21 urchin sperm (Falugi et al., 2008a), as well as in the eggs (Angelini et al., 2003), where it has an
22
23 important role in the fertilisation process (Falugi et al., 2008a). We believe that when these
24
25 carbon-based nanomaterials become attached to the sperm surface, they will affect ACh release
26
27 from the sperm **during the fertilization process**, possibly by inhibition of the membrane-
28
29 exposed choline acetyltransferase, or even through inhibition of the nicotinic ACh receptors or
30
31 AChE at the surface of the oocyte during fertilisation. This will lead to anomalous influx of Na²⁺,
32
33 and will thus affect membrane depolarisation. These events can lead to altered ACh synthesis
34
35 through choline acetyltransferase in fertilised eggs, and will subsequently alter the expression of
36
37 muscarinic receptors (Harrison et al., 2002). It is these muscarinic receptors that are responsible
38
39 for modulation of the intracellular concentrations of Ca²⁺, which is a major second messenger
40
41 that provides key activating signals in non-parthenogenic development, through cell-to-cell
42
43 communication, cell-cycle control, protein (including cholinesterases) synthesis, and cell
44
45 locomotion (Wilding et al., 1996; Kashir et al., 2013). Consequently, the specific activities of the
46
47 cholinesterases in the embryos that show developmental anomalies differ from those in the
48
49 controls.
50
51
52
53
54
55
56
57
58
59

60 ***4.2.2. Alterations in embryonic and larval morphology***

61
62
63
64
65

1
2
3
4 As discussed in the previous section, any disturbances to the process of fertilisation can result in
5
6 the loss of downstream signal transduction, which can lead to altered intracellular ion
7
8 concentrations, including Ca^{2+} (Van Koppen and Kaiser, 2003). As reported by Ravera et al.
9
10 (2006) and Falugi et al. (2008a), ionic dynamics are crucial for correct cell migration during
11
12 gastrulation and embryonic development of sea urchins. This process is again ultimately
13
14 dependent on ACh availability.
15
16

17
18 Our data here show that egg fertilisation with the CB-exposed sperm influences cell-cell adhesion
19
20 during gastrulation, which includes altered directional migration and ingression of primary
21
22 mesenchymal cells (Malinda et al., 1994; Miller et al., 1997). Gambardella et al. (2013) did not
23
24 observe these changes in a same experimental set-up where the sperm were exposed to Ag, Co or
25
26 TiO_2 nanomaterials. Again, we believe that the lower adsorption potential of the nanomaterials
27
28 investigated by Gambardella et al. (2013) resulted in their lower adsorption onto the sperm cells,
29
30 and consequently in lower levels of disturbance during egg fertilisation and the early
31
32 developmental events.
33
34
35
36

37
38 Furthermore, we show here that the exposure of sea urchin sperm to both of the carbon-
39
40 based nanomaterials tested results in anomalous arrangements of the skeletal rods. These effects
41
42 were even more pronounced at the developmental stages of plutei (48 h, 72 h, after fertilisation),
43
44 and they might also be a consequence of impaired primary mesenchymal cell migration, which is
45
46 responsible for larval skeletogenesis (Peterson and McClay, 2003; Bradham et al., 2004;
47
48 Kominami and Takata, 2004; Gambardella et al., 2015). Altered skeletogenesis as a common
49
50 response to environmental stress, and this was also reported after exposure of sea urchin gametes
51
52 (Pesando et al., 2003; Gambardella et al., 2013) and early embryos (Carballeira et al., 2012; Siller
53
54 et al., 2013) to other environmental contaminants. On the other hand, Manno et al. (2012)
55
56 reported altered biomineralisation processes due to carbon nanoparticle exposure of pluteus
57
58
59
60
61
62
63
64
65

1
2
3
4 larvae, through the activation of one of the genes controlling skeletogenesis. They explained that
5
6 the larvae activate a defence process against the external material by incorporation of the
7
8 nanoparticles into microstructures of aragonite, similar to pearl oysters.
9

10
11 Previous studies (Angelini et al., 2003; Ravera et al., 2006; Falugi et al., 2008a; Aluigi et
12
13 al., 2010) have reported that ACh availability is ultimately important for ionic dynamics, and
14
15 consequently for correct cell migration during gastrulation and embryonic development in sea
16
17 urchins. The data from the present study also show reduced activities of the cholinesterases. This
18
19 thus provides experimental confirmation that developmental anomalies are accompanied by
20
21 alterations in the activity of the cholinergic system, and it can be used as a reliable and sensitive
22
23 biomarker of developmental anomalies.
24
25
26
27
28
29
30

31 **5. Conclusions**

32
33 Here we have described the consequences of exposure of *P. lividus* sperm to the carbon-based
34
35 nanomaterials CB and GO, in terms of the egg fertilisation and early developmental stages of the
36
37 purple sea urchin (i.e., gastrulae, plutei).
38
39

40 We have confirmed the more severe effects of CB on egg fertilisation by these sperm, while the
41
42 exposure to GO had no effects on egg fertilisation up to a sperm-exposure concentration of 1
43
44 mg/L. However, embryos and larvae obtained after egg fertilisation with the *P. lividus* sperm
45
46 exposed to either of these carbon-based nanomaterials showed similar morphological effects and
47
48 inhibition of the activities of the enzymes tested: AChE and PChE. We believe that these effects
49
50 are a consequence of the attachment of these carbon-based nanomaterials to the surface of the
51
52 sperm, and their consequent interactions with fertilisation events. Carbon-based nanomaterials
53
54 can disrupt the cholinergic system of the gamete cell surface, which will lead to alterations in the
55
56 signalling pathways involved in early embryonic development. In particular, sperm exposure to
57
58
59
60
61
62
63
64
65

1
2
3
4 very low amounts of these carbon-based nanomaterials can provoke these above-described effects
5
6 on the developmental stages of *P. lividus*. **To date no environmental concentration in**
7
8 **seawater is available either in bibliography or in official databases for CB or GO. Anyway,**
9
10 **the predicted environmental concentration reported for other substances such as**
11
12 **benzo[a]pyrene (CAS No 50-32-8), involved in carbon black production process, is about**
13
14 **0.015 µg/L (https://www.env.go.jp/en/chemi/chemicals/profile_esrac/profile5/pf1-22.pdf),**
15
16 **representing a lower concentration than those tested in the present study.** However, it is
17
18 worth noting that the reported environmental concentrations of CB in polluted freshwaters (as
19
20 0.08 µg/mL to 7.5 µg/mL, CAS RN, 2013), correspond to the concentrations that are shown to
21
22 have toxic effects on sea urchins in the present study. **In conclusion, we can speculate that the**
23
24 **adverse effects of carbon-based NMs on sperm and embryos of the sea urchin might occur**
25
26 **also on other shallow water species, in particular those forming the benthonic communities,**
27
28 **where the more or less agglomerated nanomaterials sink and accumulate.**
29
30
31
32
33
34
35
36
37

38 **Acknowledgments**

39
40 The authors would like to gratefully acknowledge RITMARE Flagship Project, a National
41
42 Research Programme funded by the Italian Ministry of University and Research (MIUR) and the
43
44 Slovene Human Resources Development and Scholarship Fund, for financial support, of Tina
45
46 Mesarič. A special thank to Dr. Christopher Berrie for reading and critical appraisal of the
47
48 manuscript.
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

1
2
3
4 **References**
5

6 **Alberts, B., Bray, D., Lewis, J., Raff, M., Roberts, K., Watson, J.D. 1989. Molecular biology**
7 **of the cell. Garland Publishing, New York.**
8

9
10
11 Aluigi, M.G., Falugi, C., Mugno, M.G., Privitera, D., Chiantore, M. 2010. Dose-dependent
12 effects of chlorpyrifos, an organophosphate pesticide, on metamorphosis of the sea urchin,
13 *Paracentrotus lividus*. *Ecotoxicology* 19, 520-529.
14
15

16
17
18 Amemiya, S. 1996. Complete regulation of development throughout metamorphosis of sea urchin
19 embryos devoid of macromeres. *Dev. Growth Differ.* 38, 465-476.
20

21
22
23 American Society for Testing Materials (ASTM). 2004. Standard guide for conducting static
24 acute toxicity tests with embryos of echinoid embryos. E1563-E1598.
25

26
27
28 Angelini, C., Amaroli, A., Falugi, C., Di Bella, G., Matranga, V. 2003. Acetylcholinesterase
29 activity is affected by stress conditions in *Paracentrotus lividus* coelomocytes. *Mar. Biol.*
30 143, 623-628.
31
32

33
34
35 **Angelini, C., Baccetti, B., Piomboni, P., Trombino, S., Aluigi, M.G., Stringara, S., Gallus,**
36 **L., Falugi, C. 2004. Acetylcholine synthesis and possible functions during sea urchin**
37 **development. Eur. J. Histochem. 49, 235-244.**
38

39
40
41 **Arizzi Novelli, A., Volpi Ghirardini, A., Ghetti, P.F. 2001. Sperm cell and embryotoxicity**
42 **test with *Paracentrotus lividus* LMK. Reproducibility and comparison of sensitivity**
43 **with other echinoid species using copper. Biol. Mar. Medit. 8.**
44

45
46
47 **Arizzi Novelli, A., Argese, E., Tagliapietra, D., Bettiol, C., Volpi Ghirardini, A., 2002.**
48 **Toxicity of tributyltin and triphenyltin to early life-stages of *Paracentrotus lividus***
49 **(Echinodermata: Echinoidea). Environ. Toxicol. Chem. 21, 859-864.**
50

51
52
53 **Arizzi Novelli, A., Losso C., Falugi C., Giuliani S., Kozinkova L., Lera S., Leoni T., Manzo**
54 **S., Mazziotti C., Pellegrini D., Picone M., Volpi Ghirardini, A. 2007 a. Il test di**
55
56
57

1
2
3
4 **fecondazione con il riccio di mare *Paracentrotus lividus* (LMK) Biol. Mar. Medit. 14,**
5
6 **43-47.**

7
8
9 **Arizzi Novelli, A., Losso C., Volpi Ghirardini A., Ghetti, P.F. 2007 b. Saggi di tossicità con**
10
11 **il riccio di mare *Paracentrotus lividus*: percorso di validazione di metodologie per gli**
12
13 **ambienti di transizione attraverso una procedura di controllo qualità. Biol. Mar.**
14
15 **Medit. 14:100-102.**

16
17
18
19 **ASTM (American Society for Testing, Materials), 2004. Standard guide for conducting**
20
21 **static acute toxicity tests with embryos of echinoid embryos. E1563–E1598.**

22
23
24 Ates, M., Daniels, J., Arslan, Z., Farah, O.I., Rivera, H.F. 2013. Comparative evaluation of
25
26 impact of Zn and ZnO nanoparticles on brine shrimp (*Artemia salina*) larvae: effects of
27
28 particle size and solubility on toxicity. Environ. Sci. Processes Impacts 15, 225.

29
30
31 Baccetti, B., Burrini, A.G., Collodel, G., Falugi, C., Moretti, E., Piomboni, P. 1995. Localization
32
33 of two classes of acetylcholine receptor-like molecules in sperms of different animal species.
34
35 *Zygote* 3, 207-217.

36
37
38 **Boudouresque, C.F., Verlaque, M., 2001. Ecology of *Paracentrotus lividus*. In: Lawrence,**
39
40 **J.M. (Ed.) Edible sea urchins: biology and ecology 32. Elsevier Science, Amsterdam**
41
42 **177-216.**

43
44
45 Bradham, C.A., Miranda, E.L., McClay, D.R. 2004. PI3K inhibitors block skeletogenesis but not
46
47 patterning in sea urchin embryos. Dev. Dynam. 229, 713-721.

48
49
50 Buznikov, G.A., Shmukler, Y.B., Lauder, H.M. 1996. From oocyte to neuron: do
51
52 neurotransmitters function in the same way through development? Cell. Mol. Neurobiol. 16,
53
54 533-559.

1
2
3
4 Canesi, L., Ciacci, C., Betti, M., Fabbri, R., Canonico, B., Fantinati, A., Marcomini, A., Pojana,
5
6 G. 2008. Immunotoxicity of carbon black nanoparticles to blue mussel hemocytes. Environ.
7
8 Int. 34, 1114-1119.
9

10
11 Canesi, L., Ciacci, C., Vallotto, D., Gallo, G., Marcomini, A., Pojana, G. 2010a. *In-vitro* effects
12
13 of suspensions of selected nanoparticles (C60 fullerene, TiO₂, SiO₂) on *Mytilus* hemocytes.
14
15 Aquat. Toxicol. 96, 151-158.
16
17

18
19 Canesi, L., Fabbri R., Gallo, G., Vallotto, D. 2010b. Biomarkers in *Mytilus galloprovincialis*
20
21 exposed to suspensions of selected nanoparticles (Nano carbon black, C-60 fullerene, nano-
22
23 TiO₂, nano-SiO₂). Aquat. Toxicol. 100, 168-177.
24
25

26
27 Carata, E., Tenuzzo, B.A., Arnò, F., Buccolieri, A., Serra, A., Manno, D., Dini, L. 2012. Stress
28
29 response induced by carbon nanoparticles in *Paracentrotus lividus*. Int. J. Mol. Cell. Med. 1,
30
31 30-38.
32

33
34 Carballeira, C., Ramos-Gómez, J., Martín-Díaz, L., DelValls, T.A. 2012. Identification of
35
36 specific malformations of sea urchin larvae for toxicity assessment: application to marine
37
38 pisciculture effluents. Mar. Environ. Res. 77, 12-22.
39

40
41 CAS RN 2013, Screening assessment for the challenge: Carbon Black. 2013. Chemical Abstracts
42
43 Service Registry: 1333-86-4.
44

45 **Drews, U., 1975. Cholinesterase in embryonic development. Prog. Histochem. Cytochem. 7,**
46
47 **1-52.**
48
49

50
51 Ellman, G.L., Courtney, K.D., Andres, V.J., Featherstone R.M. 1961. A new rapid colorimetric
52
53 determination of acetylcholinesterase activity. Biochem. Pharmacol. 7, 88-95.
54

55 **Falugi, C., Prestipino, G., 1987. Effects of some inhibitors of the cholinergic system, active**
56
57 **during the sea urchin *Paracentrotus lividus* development. In: Boudouresque, C.F. (Ed.),**
58
59
60
61

1
2
3
4 **Colloque International sur *Paracentrotus lividus*. Gis Posidonie, Marseille, France, pp.**
5
6 **147-155.**
7

8
9 Falugi, C., Pieroni, M., Moretti, E. 1993. Cholinergic molecules and sperm functions. J. Submicr.
10
11 Cytol. Path. 25, 63-69.
12

13
14 Falugi, C., Lammerding-Koppel, M., Aluigi, M.G. 2008a. Sea urchin development: an alternative
15
16 model for mechanistic understanding of neurodevelopment and neurotoxicity. Birth Defects
17
18 Embryo Res. C Embryo Today 84, 188-203.
19

20
21 **Falugi, C., Collodel, G., Moretti, E., Piomboni, P., Aluigi, M.G., Angelini, C., Baccetti, B.,**
22
23 **2008b. A neural-like model for sperm-egg interactions. Involvement of the cholinergic**
24
25 **system. In: Collodel, G., Moretti, E. (Eds), Sperm Morphology and Pathology, 2008:**
26
27 **000-000 ISBN: 978-81-308-0280-0.**
28
29

30
31 Falugi, C., Aluigi, M.G., Chiantore, M.C., Privitera, D., Ramoino, P., Gatti, M.A., Fabrizi, A.,
32
33 Pinsino, A., Matranga, V. 2012. Toxicity of metal oxide nanoparticles in immune cells of the
34
35 sea urchin. Mar. Environ. Res. 76, 114-121.
36
37

38
39 Falugi, C., Aluigi, M.G. 2012. Early appearance and possible functions of non-neuromuscular
40
41 cholinesterase activities. Front. Mol. Neurosci. 5, 1-12.
42

43
44 Gambardella, C., Aluigi, G. M., Ferrando, S., Gallus, L., Ramoino, P., Gatti, A.M., Rottigni, M.,
45
46 Falugi, C. 2013. Developmental abnormalities and changes in cholinesterase activity in sea
47
48 urchin embryos and larvae from sperm exposed to engineered nanoparticles. Aquat. Toxicol.
49
50 130-131, 77-85.
51

52
53 Gambardella, C., Mesarič, T., Milivojević, T., Sepčić, K., Gallus, L., Carbone, S., Ferrando, S.,
54
55 Faimali, M. 2014. Effects of selected metal oxide nanoparticles on *Artemia salina* larvae:
56
57 evaluation of mortality and behavioural and biochemical responses. Environ. Monit. Assess.
58
59 186, 4249-4259.
60
61

- 1
2
3
4 Gambardella, C., Ferrando, S., Morgana, S., Gallus, L., Ramoino, P., Ravera, S., Bramini, M.,
5
6 Diaspro A., Faimali M., Falugi C. 2015. Exposure of *Paracentrotus lividus* male gametes to
7
8 engineered nanoparticles affects skeletal biomineralization and larval plasticity. *Aquat.*
9
10 *Toxicol.*, 158, 181-191.
11
12
13
14 Ghirardini, A.V., Novelli, A.A., Losso, C., Ghetti, P.E. 2005. Sperm cell and embryo toxicity
15
16 tests using the sea urchin *Paracentrotus lividus* (LmK). *Techniq. Aquat. Toxicol.* 8, 2.
17
18
19 **Gibbons, I.R., 1981. Cilia and flagella of eukaryotes. J. Cell Biol. 91, 107s-124s.**
20
21 **Giudice, G., 1973. Developmental biology of the sea urchin embryo. Academic**
22
23 **Press, USA. pp. 469.**
24
25
26 Goldberg, E.D. *Black Carbon in the Environment: Properties and Distribution.* John Wiley &
27
28 Sons, New York (1985) 198.
29
30
31 Harrison, P.K., Falugi, C., Angelini, C., Whitaker, M. 2002. Muscarinic signalling affects
32
33 intracellular calcium concentration during the first cell cycle of sea urchin embryos. *Cell*
34
35 *Calcium* 31, 289-297.
36
37
38 **Hull, M.S., Kennedy, A.J., Steevens, J.A., Bednar, A.J., Weiss, J.C.A., Vikesland, P.J. 2009.**
39
40 **Release of metal impurities from carbon nanomaterials influences aquatic toxicity.**
41
42 ***Environ. Sci. Technol.* 43, 4169-4174.**
43
44
45
46 Hussain, S., Boland, S., Baeza-Squiban, A., Hamel, R., Thomassen, C.L., Martens, A.J., Billon-
47
48 Galland, M.A., Fleury-Feith, J., Mosain, F., Paireon, J.C., Marano, F. 2009. Oxidative stress
49
50 and proinflammatory effects of carbon black and titanium dioxide nanoparticles: role of
51
52 particle surface area and internalized amount. *Toxicology* 260, 142-149.
53
54
55
56 Kashir, J., Deguchi, R., Jones, C., Coward, K., Stricker, S.A. 2013. Comparative biology of
57
58 sperm factors and fertilization-induced calcium signals across the animal kingdom. *Mol.*
59
60 *Reprod. Dev.* 80, 787-815.
61
62
63
64
65

- 1
2
3
4 Klaine, S.J., Alvarez, P.J., Batley, G.E., Fernandes, T.F., Handy, R.D., Lyon, D.Y., Mahendra, S.,
5
6
7 McLaughlin, M.J., Lead, J.R. 2008. Nanomaterials in the environment: behaviour, fate,
8
9 bioavailability and effects. *Environ. Toxicol. Chem.* 27, 1825-1851.
- 10
11 Kominami, T., Takata, H. 2004. Gastrulation in the sea urchin embryo: a model system for
12
13 analysing the morphogenesis of a monolayered epithelium. *Dev. Growth Differ.* 46, 309-326.
- 14
15
16 Malinda, K.M., Ettensohn, C.A. 1994. Primary mesenchyme cell migration in the sea urchin
17
18 embryo: distribution of directional cues. *Dev. Biol.* 164, 562-578.
- 19
20
21 Manno, D., Carata, E., Tenuzzo, B., Panzarini, E., Buccolieri, A., Filippo, A., Rossi, M., Serra,
22
23 A., Dini, L. 2012. High ordered biomineralization induced by carbon nanoparticles in the sea
24
25 urchin *Paracentrotus lividus*. *Nanotechnology* 23, 495104.
- 26
27
28 Manno, D., Serra, A., Buccolieri, A., Panzarini, E., Carata, E., Tenuzzo, B., Izzo, D., Vergallo,
29
30 C., Rossi, M., Dini, L. 2013. Silver and carbon nanoparticles toxicity in sea urchin
31
32 *Paracentrotus lividus* embryos. *BioNanoMaterials* 14, 229-238.
- 33
34
35
36 **Manzo, S., Buono, S., Cremisini, C. 2008. Predictability of copper, irgarol, and diuron**
37
38 **combined effects on sea urchin *Paracentrotus lividus*. *Arch. Environ. Contam. Toxicol.***
39
40 **54, 57-68.**
- 41
42
43 Manzo, S., Miglietta, M.L., Rametta, G., Buono, S., Di Francia, G. 2013. Embriotoxicity and
44
45 spermioxicity of nanosized ZnO for Mediterranean sea urchin *Paracentrotus lividus*. *J.*
46
47 *Hazard. Mater.* 254-255, 1-9.
- 48
49
50 Masiello, C.A., Druffel, E.R.M. 1998. Black carbon in deep-sea sediments. *Science* 280, 1911-
51
52 1913.
- 53
54
55 Matranga, V., Corsi, I. 2012. Toxic effects of engineered nanoparticles in the marine
56
57 environment: model organism and molecular approaches. *Mar. Environ. Res.* 76, 32-40.
- 58
59
60
61
62
63
64
65

1
2
3
4 Mesarič, T., Baweja, L., Drašler, B., Drobne, D., Makovec, D., Dušak, P., Dhawan, A., Sepčić,
5
6 K. 2013a. Effects of surface curvature and surface characteristics of carbon-based
7
8 nanomaterials on the adsorption and activity of acetylcholinesterase. Carbon 62, 222-232.
9

10
11 Mesarič, T., Sepčić, K., Piazza, V., Gambardella, C., Garaventa, F., Drobne, D., Faimali, M.
12
13 2013b. Effects of nano carbon black and single-layer graphene oxide on settlement, survival
14
15 and swimming behaviour of *Amphibalanus amphitrite* larvae. Chem. Ecol. 29, 643-652.
16
17

18
19 **Mesarič, T., Gambardella, C., Milivojević, T., Faimali, M., Drobne, D., Falugi, C.,**
20
21 **Makovec, D., Jemec, A., Sepčić, K. 2015. High surface adsorption properties of carbon-**
22
23 **based nanomaterials are responsible for mortality, swimming inhibition, and**
24
25 **biochemical responses in *Artemia salina* larvae. J. Aquat. Toxicol., in press.**
26
27 **[doi:10.1016/j.aquatox.2015.03.014](https://doi.org/10.1016/j.aquatox.2015.03.014)**
28
29

30
31 Middleburg, J.J., Nieuwenhuize, J., Can Breuge, P. 1999. Black carbon in marine sediments.
32
33 Mar. Chem. 65, 245-252.
34

35
36 Miglietta, M.L., Rametta, G., Francia, G.D., Manzo, S., Rocco, A., Carotenuto, R., Picione,
37
38 L.D.F., Buono, S. 2011. Characterization of nanoparticles in seawater for toxicity assessment
39
40 towards aquatic organisms. Sensor. Microsy. 91, 425-429.
41
42

43
44 Miller, J.R., McClay, D.R. 1997. Characterization of the role of cadherin in regulating cell
45
46 adhesion during sea urchin development. Dev. Biol. 15, 323-339.
47

48
49 Miller, R.J., Bennett, S., Keller, A.A., Pease, S., Lenihan, H.S. 2012. TiO₂ nanoparticles are
50
51 phototoxic to marine phytoplankton, PLoS ONE 7, e30321.
52

53 **Ministry of the Environment, Government of Japan, Profiles of the Initial Environmental**
54
55 **Risk Assessment of Chemicals, Vol. 5,**
56
57 **https://www.env.go.jp/en/chemi/chemicals/profile_esrac/profile5/pf1-22.pdf, accessed**
58
59 **26.3.2015.**
60
61
62
63
64
65

- 1
2
3
4 Nielsen, H.D., Berry, L.S., Stone, V., Burridge, T.R., Fernandes, F. 2008. Interactions between
5
6 carbon black nanoparticles and the brown algae *Fucus serratus*: inhibition of fertilization and
7
8 zygotic development. *Nanotoxicology* 2, 88-97.
9
- 10
11 Peterson, R.E., McClay, D.R. 2003. Primary mesenchyme cell patterning during the early stages
12
13 following ingression. *Dev. Biol.* 254, 68-78.
14
15
- 16 Pesando, D., Huitorel, P., Dolcini, V., Angelini, C., Guidetti, P., Falugi, C. 2003. Biological
17
18 targets of neurotoxic pesticides analysed by alteration of developmental events in the
19
20 Mediterranean sea urchin, *Paracentrotus lividus*. *Mar. Environ. Res.* 55, 39-57.
21
22
- 23 Rajasree, R.S.R., Vumar, G.V., Abraham, S.L., Indabakandan, D. 2010. Studies on the
24
25 toxicological effects of engineered nanoparticles in environment – a review, *Int. J.*
26
27 *Appl. Bioengineer.* 4, 2.
28
29
- 30 Rao, C.N.R., Biswas, K.S., Subrahmanyam, K.S., Govindaraj, A. 2009. Graphene, the new
31
32 carbon. *J. Mater. Chem.* 19, 2457-2469.
33
34
- 35 Ravera, S., Falugi, C., Calzia, D., Pepe, I.M., Panfoli, I., Morelli, A. 2006. First cell cycles of sea
36
37 urchin *Paracentrotus lividus* are dramatically impaired by exposure to extremely low-
38
39 frequency electromagnetic field. *Biol. Reprod.* 75, 948-953.
40
41
- 42 Siller L., Lemloh M.L., Piticharoenphun S., Mendis B.G., Horrocks B.R., Brümmer F., Medaković
43
44 D. 2013. Silver nanoparticle toxicity in sea urchin *Paracentrotus lividus*. *Environ. Pollut.* 178,
45
46 498-502.
47
48
49
- 50 Sorahan, T., Harrington, J.M. 2007. A "lugged" analysis of lung cancer risks in UK carbon black
51
52 production workers, 1951-2004. *Amer. J. Ind. Med.* 50, 555-564.
53
54
- 55 **Tash, J.S., Kim, S., Schuber, M., Seibt, D., Kinsey, W.H. 2001. Fertilization of sea urchin**
56
57 **eggs and sperm motility are negatively impacted under low hypergravitational forces**
58
59 **significant to space flight. *Biol. Reprod.* 65, 1224-1231.**
60
61
62
63
64
65

1
2
3
4 Tyrode, M.V. 1910. The mode of action of some purgative salts. Arch. Int. Pharmacodyn. 20,
5
6 205.

7
8
9 Van Koppen, C.J., Kaiser, B. 2003. Regulation of muscarinic acetylcholine receptor signalling.
10
11 Pharmacol. Ther. 98, 197-220.

12
13
14 **Volpi Ghirardini, A., Arizzi Novelli A. 2002. A cell sperm toxicity test procedure for the**
15
16 **Mediterranean species *Paracentrotus lividus* LMK (Echinodermata: echinoidea).**
17
18 **Environ. Technol. 22, 439-445.**

19
20
21 Wilding, M., Wright, E.M., Patel, R., Ellis-Davies, G., Whitaker, M. 1996. Local perinuclear
22
23 calcium signals associated with mitosis entry in early sea urchin embryos. J. Cell Biol. 135,
24
25 191-199.

26
27
28 Wang, Y., Li, Z., Wang, J., Li, J., Lin, Y. 2011. Graphene and graphene oxide:
29
30 biofunctionalization and applications in biotechnology. Trends Biotechnol. 29, 205-212.

31
32
33 Xia, X.R., Monteiro-Riviere, N.A., Mathur, S., Song, X., Xiao, L., Oldenberg, S.J., Fadeel, B.,
34
35 Riviere, J.E. 2011. Mapping the surface adsorption forces of nanomaterials in biological
36
37 systems, ACS Nano. 5, 9074-9081.

38
39
40 **Zhang, X.J., Yang, L., Zhao, Q., Caen, J.P., He, H.Y., Jin, Q.H., Guo, L.H., Alemany, M.,**
41
42 **Zhang, L.Y., Shi, Y.F. 2002. Induction of acetyl-cholinesterase expression during**
43
44 **apoptosis in various cell types. Cell Death Diff. 9, 790-800.**
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

1
2
3
4 **Figure legends:**
5
6
7
8

9 **Figure 1.** Egg fertilisation by the *P. lividus* sperm exposed for 1 h to CB (**A**) and GO (**B**),
10 according to the exposure concentrations, as compared to the controls (0). Data are means
11 \pm standard error of three independent experiments.
12
13
14

15
16
17
18 **Figure 2. Larval development** anomalies and their classification following egg fertilisation by
19 the *P. lividus* sperm exposed for 1 h to CB (**A, B**) and GO (**C, D**), according to the 24-h-old
20 gastrulae (**A, C**) and 72-h-old plutei (**B, D**) and the exposure concentrations, as compared to the
21 controls (0). ND, non-developed gastrulae or arrested development; AD, anomalous
22 development; D, normal development; del, delayed forms; sk, skeletal anomalies. Data are means
23 \pm standard error of three independent experiments.
24
25
26
27
28
29
30
31

32
33
34
35 **Figure 3.** Representative morphological features of embryos 24 h after egg fertilisation (gastrula
36 stage) by the *P. lividus* sperm exposed to CB or GO (as indicated). Black asterisks, anomalous
37 migration of primary mesenchymal cells; white asterisks, loss of cell-cell adhesion; arrow,
38 absence of skeleton in gastrula; arrowheads, gastrulae with irregular shapes. Scale bars: 50 μ m.
39
40
41
42
43
44
45
46
47

48 **Figure 4.** Representative morphological features of embryos 48 h (**A**) and 72 h (**B**) after egg
49 fertilisation (pluteus stages) by the *P. lividus* sperm exposed to CB or GO (as indicated).
50 Asterisks, incomplete or erroneous migration of skeleton elements; arrows, anomalous
51 arrangement of skeletal rods (crossed tips). Scale bar: 50 μ m.
52
53
54
55
56
57
58
59
60
61
62
63
64
65

1
2
3
4 **Figure 5.** Acetylcholinesterase (A,B) and propionylcholinesterase activity (C, D) in 24 h old *P.*
5
6 *lividus* embryos obtained after egg fertilisation with sperm exposed to either carbon black (A,C)
7
8 or graphene oxide (B,D). Data, representing the mean value \pm SE (n = 3), were analysed by
9
10 ANOVA followed by Bonferroni *post hoc* test. * p<0.05. CB, carbon black; GO, graphene oxide.
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

1
2
3
4 **Supporting Information**
5
6
7
8

9 **Transmission electron microscopy of the nanomaterials**

10
11 For the transmission electron microscopy investigations, the nanomaterials were deposited by the
12 drying of the aqueous suspension on copper-grid-supported, perforated, transparent carbon foil.
13
14 The transmission electron microscopy analysis was performed using a JEOL 2010 F electron
15 microscope, operated at 200 kV. The carbon black (CB) is seen as agglomerates of amorphous,
16 globular primary nanoparticles, with sizes of ~20 nm (**Figure S1**). For the graphene oxide (GO),
17 according to the data from the producer, at least 80% of this is in the form of 0.5- μ m to 5- μ m-
18 wide sheet-like molecules. However, when the aqueous GO suspension was dried on the
19 transmission electron microscopy support, the GO deposited on the in the form of plate-like
20 aggregates of different thicknesses. **Figure S2** shows the edge of a thin platelet aggregate lying
21 on the perforated, amorphous-carbon-support layer. The different contrast in the image shows
22 buckling and folding of the GO layers.
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39

40 **Supplementary Figure legends**

41
42
43
44
45 **Figure S1.** Representative transmission electron microscopy image of CB.
46
47
48
49

50 **Figure S2.** Representative transmission electron microscopy image of GO.
51
52
53
54

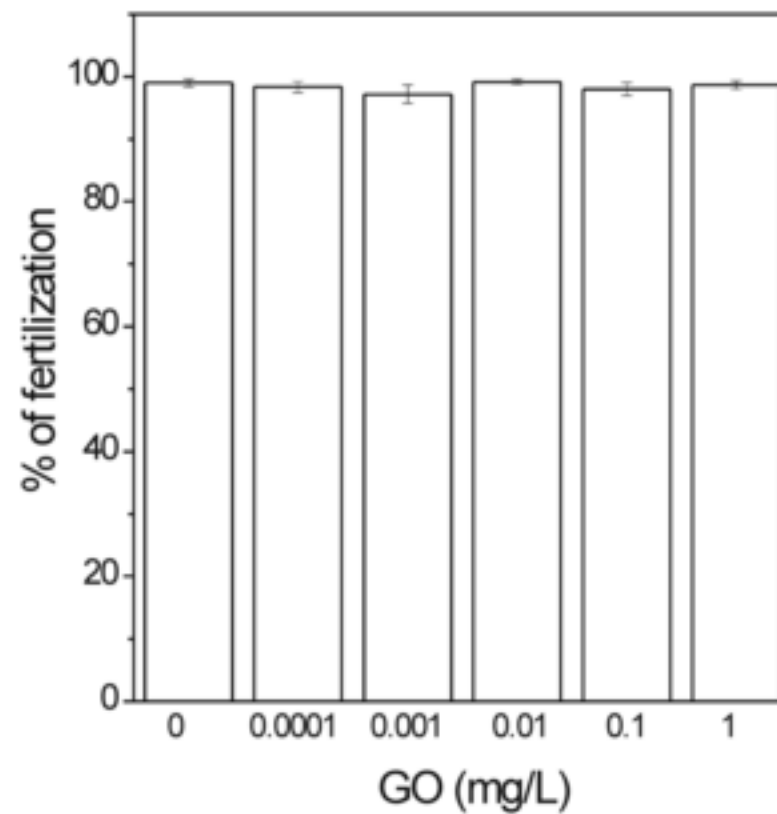
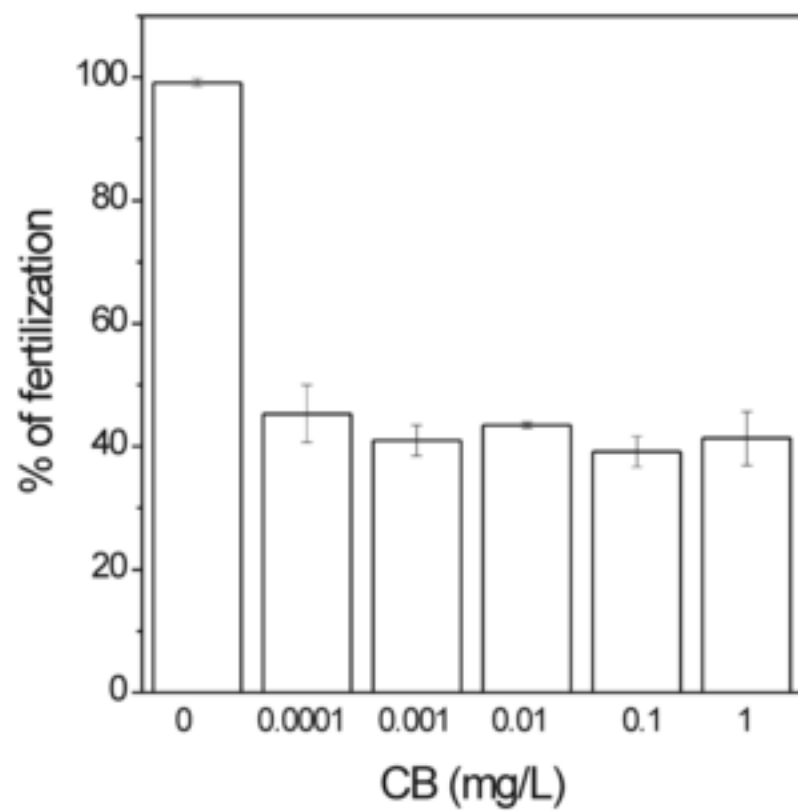
55 **Figure S3. Energy-dispersive X-ray spectroscopy (EDS) analysis of carbon-based**
56 **nanomaterials suspended in FNSW: (A) only FSNW suspensions without NM; (B) carbon**
57
58
59
60
61
62
63
64
65

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

black in FNSW, and (C) graphene oxide in FNSW. EDS spectra shown on the right correspond to position marked with white arrows. All results in tables are in weight %.

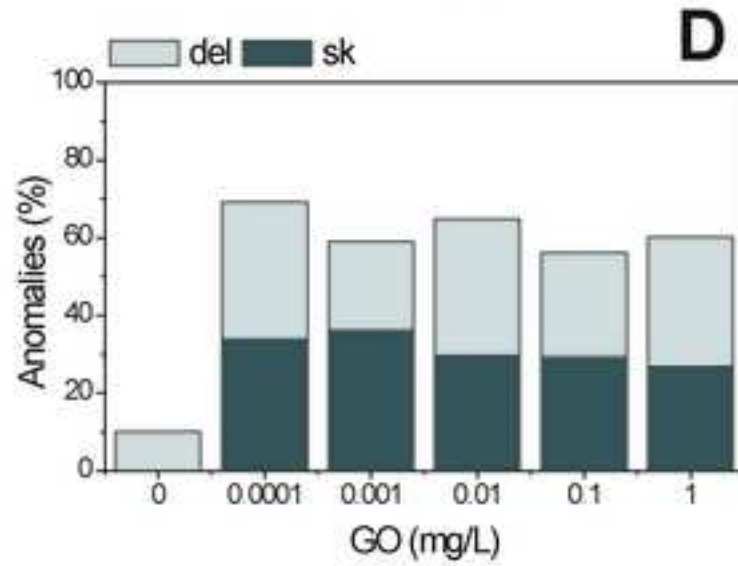
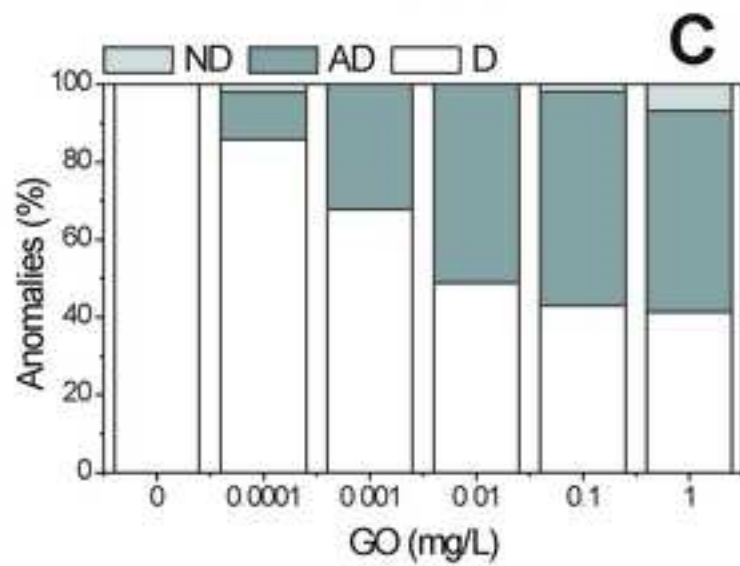
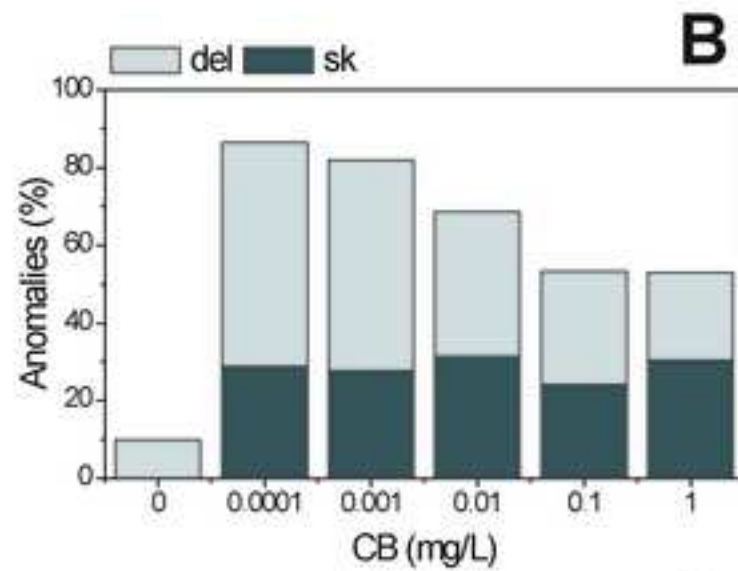
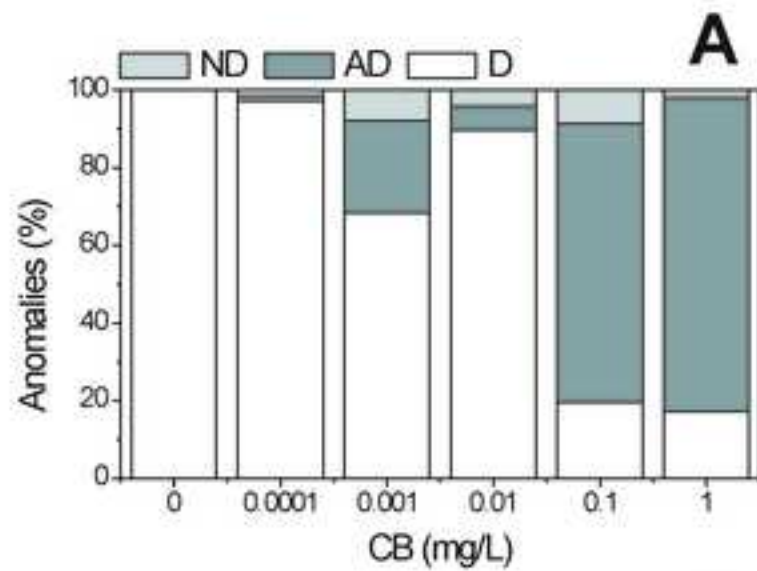
Figure(s)

[Click here to download high resolution image](#)



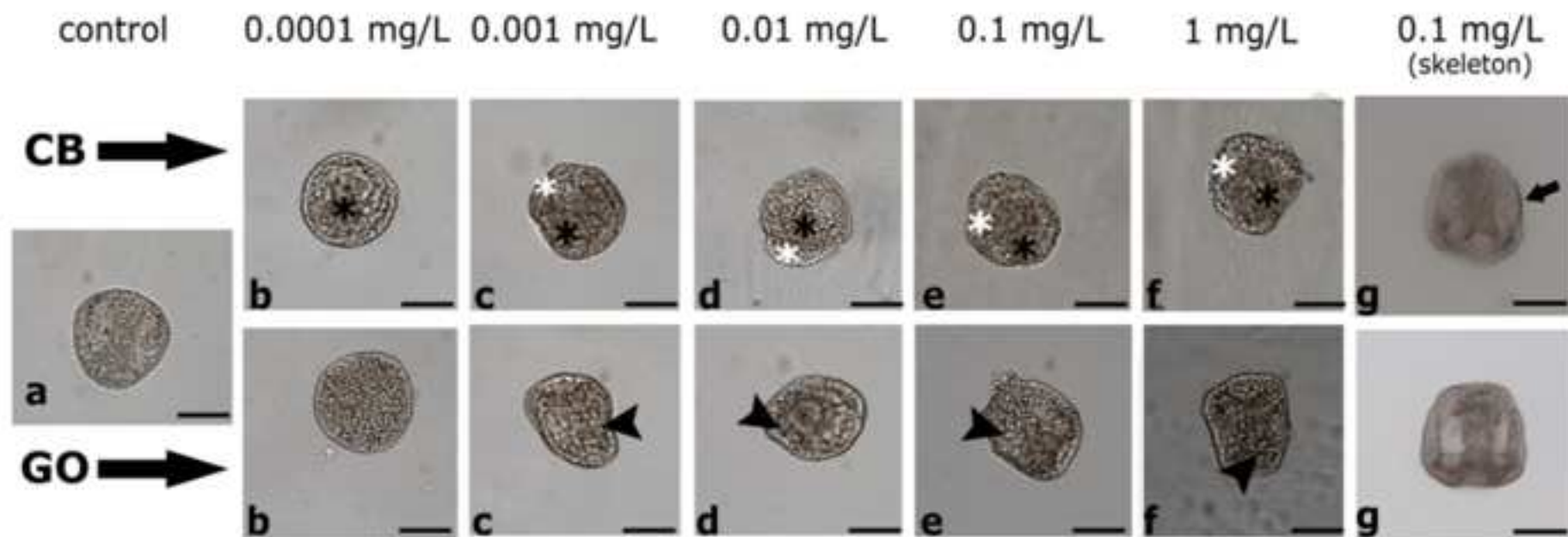
Figure(s)

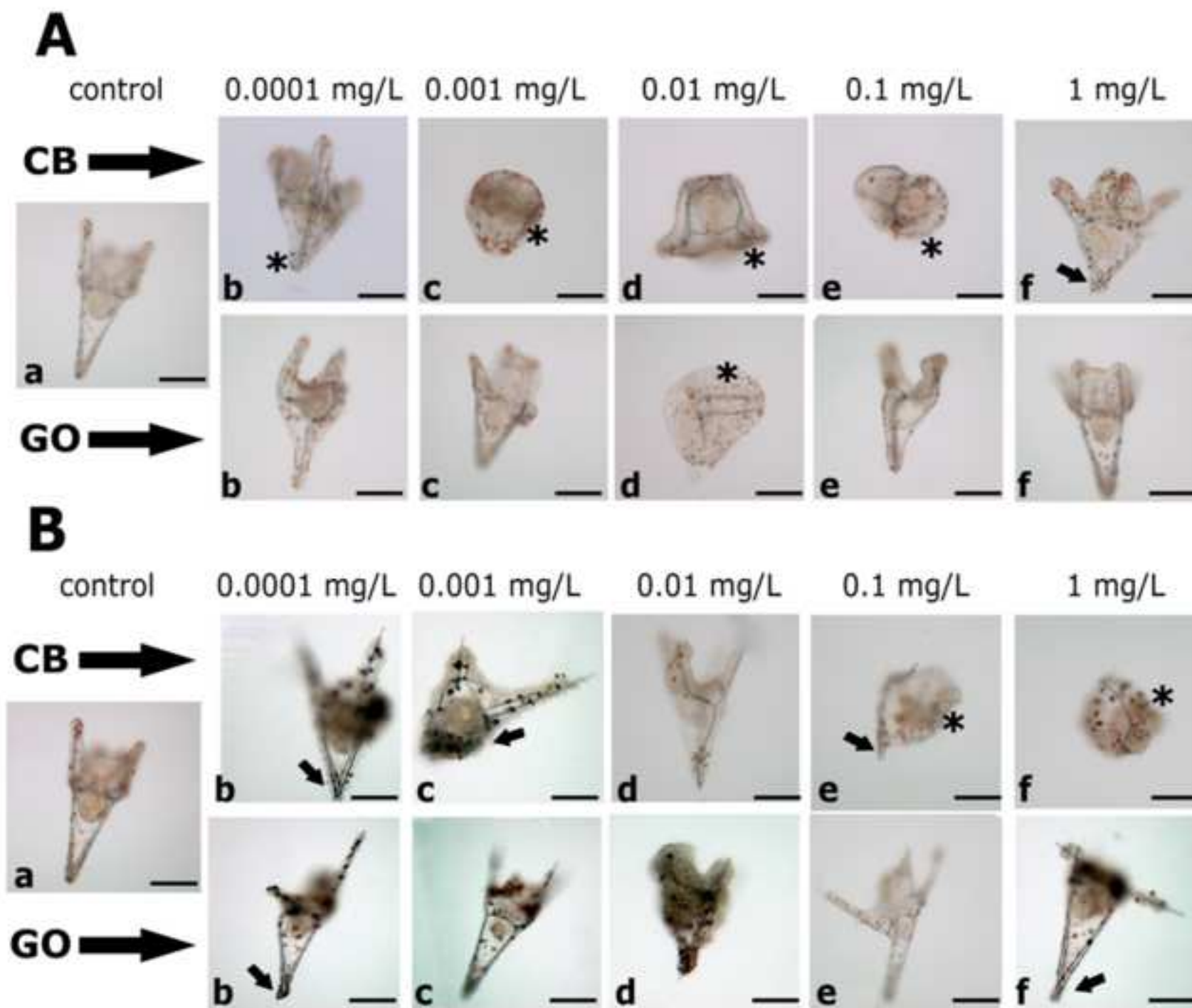
[Click here to download high resolution image](#)

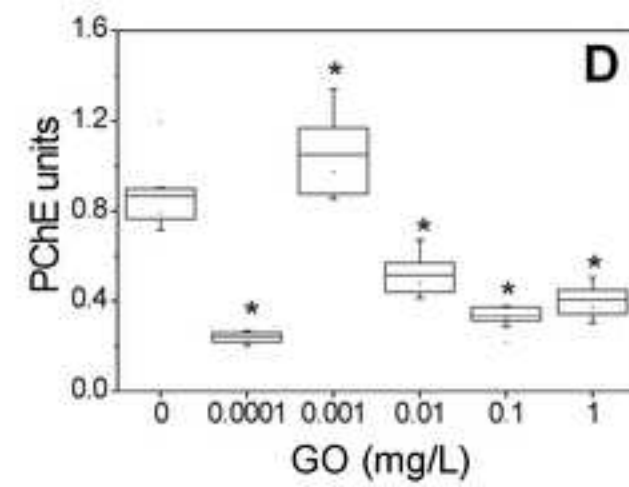
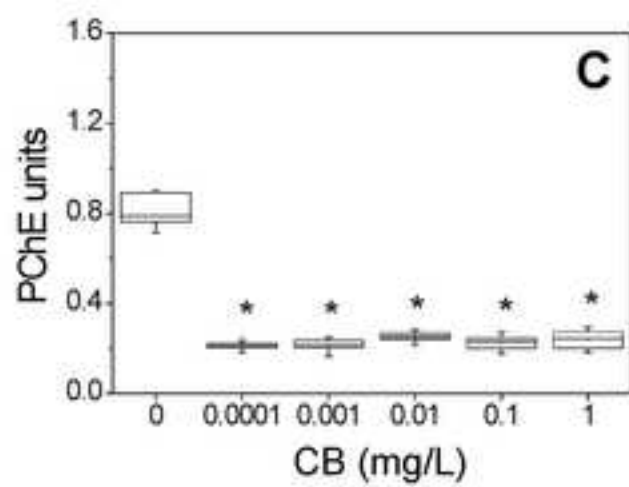
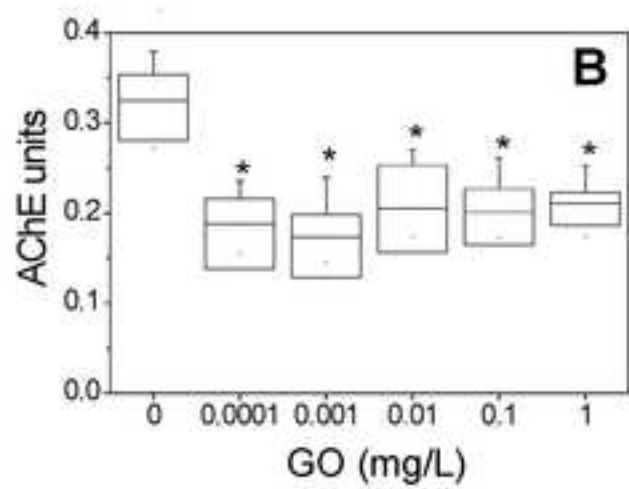
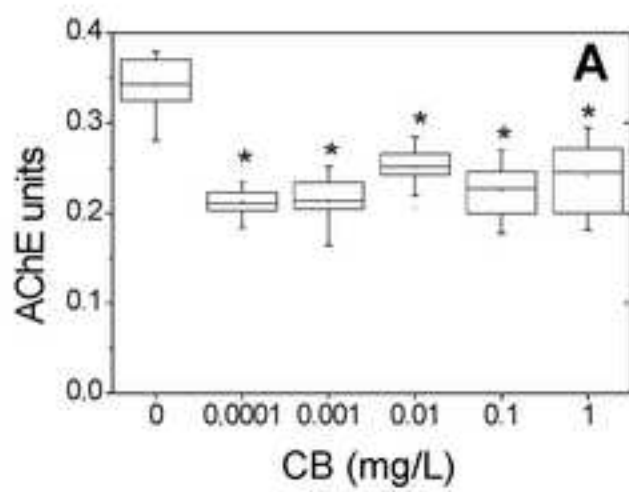


Figure(s)

[Click here to download high resolution image](#)







Supplementary Material

[Click here to download Supplementary Material: Mesari et al. Figure S1.tif](#)

Supplementary Material

[Click here to download Supplementary Material: Mesari et al. Figure S2.tif](#)

Supplementary Material

[Click here to download Supplementary Material: Mesaric et al. Figure S3.docx](#)

Highlights:

- CB exposure of *P. lividus* sperm reduces egg fertilisation already at 0.0001 mg/L
- GO exposure of *P. lividus* sperm does not affect egg fertilisation at up to 1 mg/L
- CB and GO exposure of sperm induce developmental anomalies in gastrulae and plutei
- CB and GO exposure of sperm reduce cholinesterase activities in gastrulae